

1SSN: 2476-2377

Research Article

International Journal of Cancer Research & Therapy

Assessment of Age-Dependent Effective Dose and Toxicity Risks Due To Ingestion of ²²⁶RA and ²²⁸RA in River Water Samples in a Part of Bitumen Belt, Ondo State. Nigeria

Oladele Samuel Ajayi1*, Olusegun Yemi Omogunloye2

¹Department of Physics, Federal University of Technology, P.M.B. 704, Akure. Ondo State. Nigeria.

²Department of Physical Sciences, Olusegun Agagu University of Science and Technology, Okitipupa. Ondo State. Nigeria.

*Corresponding author

Oladele Samuel Ajayi, Department of Physics, Federal University of Technology, P.M.B. 704, Akure. Ondo State. Nigeria.

Submitted: 25 Jan 2022; Accepted: 01 Feb 2022; Published: 10 Feb 2022

Citation: Oladele Samuel Ajayi, Olusegun Yemi Omogunloye. (2022). Assessment of Age-Dependent Effective Dose and Toxicity Risks Due To Ingestion of ²²⁶RA and ²²⁸RA in River Water Samples in a Part of Bitumen Belt, Ondo State. Nigeria. Int J Cancer Res Ther, 7(1), 05-13.

Abstract

Exposure to naturally occurring radioactive chemical elements in food and water is attracting the attention of many researchers in recent times because of its deleterious health effects. In this study, the activity concentrations of natural radionuclides in some river waters in southwestern Nigeria were measured using high-purity germanium (HPGe) detector. They ranged from 0.12 to 2.31, 0.17 to 2.85, and 7.86 to 65.51 Bql⁻¹ for ²²⁶Ra, ²²⁸Ra, and ⁴⁰K respectively. Mean of the total annual effective dose were 9.86, 2.46, 1.71, 2.43, 5.74, and 0.99 mSvy⁻¹ for age groups \leq 1y, 1-2y, 2-7y, 7-12y, 12-17y, and >17y respectively. Estimated mass concentration of ²²⁶Ra and ²²⁸Ra in the samples ranged from 3.29x10⁻⁶ to 59.71x10⁻⁶ µgl⁻¹ with a mean of 35.04x10⁻⁶ µgl⁻¹ and 1.69x10⁻⁸ to 28.30x10⁻⁸ µgl⁻¹ with a mean of 14.30x10⁻⁸ µgl⁻¹ respectively. Mortality and morbidity risks ranged from 0.04x10⁻³ to 0.77x10⁻³ with a mean of 0.42x10⁻³ and 0.04x10⁻³ to 0.80x10⁻³ with a mean of 0.44x10⁻³ respectively for ²²⁶Ra while they ranged from 0.11x10⁻³ to 1.89x10⁻³ with a mean of 0.96x10⁻³ and 0.16x10⁻³ to 2.66x10⁻³ with a mean of 1.38x10⁻³ respectively for ²²⁸Ra. Lifetime average daily dose (LADD) of ²²⁶Ra and ²²⁸Ra ranged from 9.39x10⁻¹⁴ µgkg⁻¹d⁻¹ to 181.01x10⁻¹⁴ µgkg⁻¹d⁻¹ with a mean of 100.06x10⁻¹⁴ µgkg⁻¹d⁻¹ and 4.82x10⁻¹⁶ µgkg-1d⁻¹ to 80.70x10⁻¹⁶ µgkg-1d⁻¹ with a mean of 40.90x10⁻¹⁶ µgkg⁻¹d⁻¹ respectively. The hazard quotient (HQ) for 226Ra and 228Ra ranged from 8.38x10⁻¹⁴ to 161.61x10⁻¹⁴ with a mean of 89.33x10⁻¹⁴ and 4.30x10⁻¹⁶ to 70.09x10⁻¹⁶ with a mean of 36.50x10⁻¹⁶ respectively. Radiological hazard indicator of radium is of concern in these river water samples.

Keywords: Activity concentration; Radiological hazard; River water; Age-dependent effective dose; Toxicity risks; Ingestion of ²²⁶Ra and ²²⁸Ra in water; Lifetime cancer risk.

Introduction

Natural radioactivity in water has been attracting widespread attention because of the health problems that radioactive materials cause when they enter the human body through drinking water. Naturally occurring radionuclides such as uranium, thorium, radium and their decay products in drinking water give rise to radiation exposure through the drinking water pathway. It has been reported that the average worldwide radiation exposure to natural sources in foods and drinking water is 0.29 mSv y⁻¹, made up of about 0.17 mSv y⁻¹ from ⁴⁰K and about 0.12 mSv y⁻¹ from uranium and thorium [1]. The World Health Organization (WHO) has recommended safe values for various drinking water quality parameters in its general guidelines, which has been used by many countries to formulate their own national water quality guidelines [2].

Radium is regarded as a highly toxic element in water. It exists in four naturally occurring isotopes ²²³Ra, ²²⁴Ra, ²²⁶Ra and ²²⁸Ra. Many studies have been conducted on the occurrence and levels of ²²⁴Ra, ²²⁶Ra and ²²⁸Ra in drinking water [3, 4, 5, 6, 7, 8, 9]. While radium- 223 is a decay product of ²³⁵U (a rare uranium isotope), ²²⁴Ra and ²²⁸Ra originate from the decay of ²³²Th. Ra-

dium -226 is a decay product of ²³⁸U. Whereas both ²²³Ra and ²²⁴Ra are very short-lived radium isotopes with half live of 11.4 days and 3.6 days respectively, ²²⁶Ra and ²²⁸Ra are more stable with half-lives of 1602 years and 5.75 years respectively and hence the predominant radium isotopes in groundwater. When disintegrating, these radium isotopes emit nuclear radiations that can penetrate and ionize matter to various levels. Although 224 Ra and 226Ra essentially emit α-radiation, which is believed to be the cause of prevailing deleterious health effects or tissue damage ²²⁸Ra emits β-radiation. Radium also releases some γ-radiation. Radium enters surface and groundwater systems through processes such as aquifer solid weathering, direct recoil over the liquid-solid limit, desorption from the sediment surfaces, etc [3, 10]. The movement of radium in water is dependent on the geochemical properties of solids in the aquifer. Its activity concentration in groundwater depends on its activity concentration in the bedrock, mechanisms like precipitation, dissolution, complexation and adsorption-desorption, which affects its transport in water. All the processes are related to the chemical composition of groundwater [11].

Due to its similar metabolism to calcium, ²²⁶Ra is a bone-seeker.

Hence it is in the skeleton that the animal ingesting ²²⁶Ra would have the greatest body burden. After ingestion, about 80 to 85% of radium is promptly expelled from the body through feces while the rest 15 to 20% enters the blood stream and is transported to all parts of the body and is deposited on bone surfaces [8]. The radium deposits on bone surfaces are eventually covered by new deposits and old deposits penetrate deeper into the bone, where it builds up in growing bones and remains in the skeleton for a long time where it can cause bone cancer. Exposure to higher levels of radium than the Maximum Contaminant Level (MCL) over a long period of time may also result in other harmful effects like anaemia, cataracts, fractured teeth, acute leucopenia, necrosis of the jaw cancers (other than bone cancer) and even death. The adverse effect depends on its amount in water (dose), quantity of water consumed, how long an individual is exposed to it (duration) and individual characteristics such as age, state of health, lifestyle etc. The U.S. Environmental Protection Agency has established MCL of 0.185 Bq l⁻¹ for combined radium (²²⁶Ra plus ²²⁸Ra) for drinking water whereas the MCL fixed by the World Health Organization is 1.0 Bq 1⁻¹ and 0.1 Bq 1⁻¹ for ²²⁶Ra and ²²⁸Ra respectively[12, 13]. Health Canada's MCL for ²²⁶Ra only is 0.5 Bq l⁻¹ [14].

Since there is no adequate supply of pipe borne or potable water in many rural communities in Nigeria, people of those communities turn to dug wells, rivers and streams (surface waters) as sources of drinking water. No water treatment plant or local water purification system is available for treating their drinking water in these rural areas to reduce or remove the microbial or radioactivity load. River water is used for direct drinking and other purposes without filtering by the local population. The area chosen for this study, with a population of about 510,700, is one of such rural areas in southwestern Nigeria. So, it is necessary to determine the activity concentrations of natural radionuclides in water of these rivers and quantitate radium in the drinking water in order to measure its dose and risk to the target demographic group. Maxwell reported a range of 0.44 to 2.7 mBq 1-1 226Ra activity concentrations in groundwater-based drinking water samples in Dawaki, Kuje, Giri and Sabon-Lugbe area of Abuja, North Central Nigeria [9]. Adekoya reported a range of 1.62 to 3.17 Bq 1⁻¹ ²²⁶Ra activity concentrations in potable drinking water samples from former tin-mining areas with elevated activity in Jos, Nigeria [15]. Agaja and Ajisafe reported a range of 0.5 to 5.6 Bq l-1 226Ra activity concentrations in borehole and surface (river) drinking water samples on coastal communities of Delta state, Nigeria [16]. In a similar investigation carried out in Tanke, Ilorin, Nigeria, Nwakwo reported ²²⁶Ra activity concentration ranging from 0.8 to 7.4 Bq l⁻¹ Available literatures on this subject shows that activity concentration of ²²⁶Ra and ²²⁸Ra in drinking water has not been determined in any part of the Bitumen belt of Ondo State, Nigeria [17]. This study was aimed at determining the natural radionuclides content of water of some selected rivers in Okitipupa and Irele areas of Ondo State, southwestern Nigeria as well as the mass concentration of ²²⁶Ra and ²²⁸Ra, and assessing the radiological implications of human exposure to ionizing radiation emitted by these radionuclides.

Materials and Methods Study Area

Ondo State lies between longitude 4°30'E and 6°0'E and latitude 6°0'N and 7°30'N with mean annual rainfall of 1,150mm in the northern areas to about 2,000mm in the southern areas; relative humidity between 70% and 85% in the southern part and less than 78% in the northern part, with mean temperature of 21°C in the south and 32°C in the north [18].

Okitipupa and Irele Local Government Areas fall within the bitumen belt of Ondo State and are predominantly populated by the Ikales of Yoruba extraction. Rivers in the area include; Ominla, Akeun, Ufara, Otu, Oha and Oni, and the vast River Oluwa. The waters of these rivers serve as a major source of drinking water and livelihood to the people of this region, since fishing is their main occupation. The people also use the waters for drinking, bathing, washing and other domestic purposes without using any water treatment methods to remove radioactivity. Bottom river sediments are also dug and used in building construction in the region.

Sample Collection and Preparation

In this study, river water samples were taken along the course of five major rivers; Akeun, Oha, Oluwa, Otu and Ufara in Irele and Okitipupa Local Government Areas of Ondo State, Southwestern Nigeria. Ten water samples were taken from the rivers, two samples per river.

The water samples were transferred into 1 litre polyethylene gallons with 1 ml of dilute hydrochloric acid added to it in order to prevent adherence of the radionuclides to the walls of the containers. Each water (with a volume of 1 litre) sample was placed in a Marinelli beaker. The Marinelli beakers were properly sealed with polyvinyl chloride (PVC) tapes to prevent escape of ²²²Rn and ²²⁰Rn from the samples. The samples in the Marinelli beakers were stored for four weeks to allow time for ²²²Rn to attain a state of secular equilibrium with its short-lived decay products prior to gamma spectroscopy. After these, the samples were taken to the Ghana Atomic Energy Commission (G.A.E.C), Legon, Accra, Ghana for analysis with the high-purity germanium (HpGe) detector.

Sample Analysis

The activity concentrations of the samples were measured by using a computerized gamma ray spectrometry system consisting of a high purity germanium (HPGe) detector of 40% relative efficiency coupled to conventional electronics, which was connected to a multichannel analyzer card (MCA) set up on a desktop. The resolution of the spectrometer was 1.8 keV for 1.33 MeV line of 60Co. User-friendly MAESTRO-30 spectra analyzer software was used to store and analyze the data, and evaluated the radionuclide activity concentrations in the water samples. The spectrometer was placed at the centre of a 5-cm thick cylindrical lead shield, which was lined with layers of 3-mm thick copper, Plexiglass and cadmium to reduce the background gamma-ray. A mixed calibration standard serial number NW 146 in form of solid water from Deutscher Kalibrierdienst (DKD-3) QSA Global GmBH, Germany was used for energy

and efficiency calibrations of the detector in the same geometry as the samples. To acquire spectral for each sample, counting was done for 36,000 s. The gamma-ray emissions of ²¹⁴Pb (351.9 keV) and ²¹⁴Bi (609.3 keV and 1120.3 keV) were used to estimate the activity concentrations of ²²⁶Ra, for determination of activity concentrations of ²²⁸Ra, gamma-ray peaks of ²²⁸Ac (911. keV) and ²⁰⁸Tl (583 keV and 2615 keV) were used. The activity concentrations of 40K were evaluated using its single emission line situated at 1460.8 keV.

The activity concentrations (Asp) of ²²⁶Ra, ²²⁸Ra and ⁴⁰K in Bq kg⁻¹ for the water and sediment samples were determined using the expression in equation 1 [19].

where, C_{sum} is the final count of the radionuclide in a sample, P_{ϵ} is the emission probability, E is the efficiency of the detector, C_{T} is the counting time, and V is the volume of sample.

Calculation of Mass Concentration of ²²⁶Ra and ²²⁸Ra

The mass concentrations of ²²⁶Ra and ²²⁸Ra in the water samples were calculated from their activity concentrations using equation 2.



Which yields

Where A_{Bq} is the activity of 226 Ra or 228 Ra (Bq), M_{Ra} is the atomic mass of 226 Ra or 228 Ra (g mol⁻¹), N_A is Avogadro constant (6.022x10²³ mol⁻¹), and $t_{1/2}$ is the half-life of 226 Ra or 228 Ra (1602 y and 5.75 y respectively).

Annual Effective Dose Rates of the River Water Samples

The annual effective dose from ingestion of radionuclides in the water samples was estimated on the basis of the activity concentrations of the radionuclides, the volume of the water intake, which depends on the age of the person taking the water, and the age-dependent dose conversion factors for the radionuclides. The annual effective dose due to water intake was calculated using equation 4 [1].

$$E_{d} = A_{c} A_{I} C_{f} \tag{4}$$

The total annual effective dose D (mSv y-1) to an individual was established by summing contributions from all radionuclides present in the water samples i.e.

$$D = \sum_{i} (Ac)_{i} A_{i} (C_{f})_{i}$$
 (5)

where i are the radionuclides ²²⁶Ra, and ²²⁸Ra, A_c is the activity concentration of the radionuclide in the water (Bq l⁻¹), Ai is the annual intake of the drinking water (l y⁻¹) and Cf is the ingested dose conversion factor for ²²⁶Ra and ²²⁸Ra (Sv Bq⁻¹), which varies with both radionuclides under consideration and the age of individuals ingesting the radionuclides.

The intake rates and conversion factors used in this study were based on the International Commission on Radiological Protection for age groups 0-1 y, 1-2 y, 2-7 y, 7-12 y, 12-17 y, and >17 y old with annual average water intake estimated as 200, 260, 300, 350, 600 and 730 litres respectively [20].

The dose conversion factors for ingestion of natural radionuclides for members of the public, according to the ICRP are different with respect to different age groups. Table 1 shows the various dose conversion factors for different age groups for the natural radionuclides (226 Ra and 228 Ra) in the water samples analyzed.

Table 1: Dose conversion factors (Sv Bq⁻¹) for ingested natural radionuclides for the general public (ICRP, 2012).

Radionuclide	≤1 year	1-2 years	2-7 years	7-12 years	12-17 years	Above 17 years
Ra-226	4.7E-6	9.6E-7	6.2E-7	8.0E-7	1.5E-6	2.8E-7
Ra-228	3.0E-5	5.7E-6	3.4E-6	3.9E-6	5.3E-6	6.9E-7

Toxicity Radiological Risk Assessment of Radium in the Water Samples

To assess the health hazards due to the ingestion of radium in the water samples in this study, two types of radium toxicity were evaluated – radiological toxicity due to ²²⁶Ra and ²²⁸Ra as emitters of high ionizing power radiations and chemical toxicity due to their being a heavy elements.

Radiological Toxicity Assessment

The radiological toxicity of radium was estimated as the life time cancer risk (LTCR) due to ingestion of radium in the water samples using equation 6 [21, 22].

$$LTCR = A_{c}C_{s}V_{c}L_{s}$$
 (6)

where A_c is activity concentration of radium (226 Ra, and 228 Ra) in drinking water sample, Cc is Cancer risk coefficient, Vc is Volume of water consumed and Le is Life expectancy.

Cancer risk coefficients of 7.17x10⁻⁹ and 1.04x10⁻⁸ Bq⁻¹ of ²²⁶Ra, and 2.00x10⁻⁸ and 2.81x10⁻⁸ of ²²⁸Ra for mortality and morbidity respectively were taken from EPA [1, 23]. The average Nigerian life expectancy at birth is 45.5 y, that is about 16,619 days for both males and females and average adult daily consumption of water is about 21[1, 13].

Chemical Toxicity Assessment Lifetime Average Daily Dose (LADD)

The chemical toxicity (non-carcinogenic) risk due to ingestion of radium in the water samples was assessed in terms of the Lifetime Average Daily Dose (LADD). LADD is expressed as the quantity of the toxic substance taken into the body per kilogram of body mass per day.

It was calculated using equation 7 [14, 21, 24, 25, 26].



where C is the mass concentration of ^{226}Ra or ^{228}Ra (µg l¹¹), IR is water consumption rate (1 day⁻¹), ED is total exposure duration (y), EF is exposure frequency (days y-1), BW is average body mass of consumer (kg) and AT is average time, which is the life expectancy (y). The water consumption rate was set at 2 l day⁻¹. The total exposure frequency was 365.25 days and total exposure duration was 45.5 y (about 16,619 days) and average man's mass of 70 kg was used for the calculation of LADD.

Hazard Quotient

Hazard quotient of a toxic material is a measure of the extent

of damage done as a result of the ingestion of the material. In this context, the materials are ²²⁶Ra and ²²⁸Ra ingested in the investigated drinking water samples. Hazard Quotient (HQ) was calculated using [22; 25, 27].



where LADD is the Lifetime Average Daily Dose and RfD is reference dose, equal 1.12 μg kg-1 day-1 [13].

Results and discussion

Activity Concentration of Natural Radionuclides in River Water Samples

Measured activity concentrations of 226 Ra, 228 Ra and 40 K in water samples of the rivers and the total annual effective dose for the six ICRP age groups 0-1 y, 1-2 y, 2-7 y, 7-12 y, 12-17 y and > 17 y are displayed in Table 2.

Table 2: Activity concentrations of radionuclides in	the water samples and total annua	d effective dose to different age groups

River	Sample	Activity concentration (Bq l-1)			Total annual effective dose (mSv y-1)					
	location	²²⁶ Ra	²²⁸ Ra	⁴⁰ K	0-1y	1-2y	2-7y	7-12y	12-17 y	> 17y
Oluwa	Opa	0.12±0.02	0.17±0.03	7.86±1.18	1.13	0.28	0.20	0.27	0.65	0.11
	Araromi Ayeka	0.58±0.08	0.52±0.08	21.34±3.20	3.67	0.92	0.64	0.87	2.18	0.38
Akeun	Ikoya 1	2.31±0.35	1.78±0.27	52.76±7.91	12.85	3.22	2.25	3.08	7.74	1.37
	Ikoya 2	1.77±0.33	2.77±0.61	65.51±2.43	18.28	4.55	3.16	4.28	10.4	1.76
Ufara	Oloto 1	0.82±0.34	2.29±0.79	56.38±1.60	14.51	3.60	2.49	3.36	8.02	1.32
	Oloto 2	2.18±0.60	2.85±0.47	64.90±1.27	19.15	4.77	3.31	4.50	11.03	1.88
Oha	Akotogbo 1	2.03±0.30	1.25±0.19	7.95±1.19	9.41	2.36	1.65	2.28	5.80	1.05
	Akotogbo 2	0.58±0.09	0.44±0.07	10.91±1.64	3.19	0.80	0.56	0.76	1.92	0.34
Otu	Iyara 1	1.09±0.16	1.22±0.18	21.81±3.27	8.35	2.08	1.45	1.97	4.86	0.84
	Iyara 2	1.31±0.20	1.14±0.17	14.26±2.14	8.07	1.94	1.41	1.92	4.81	0.84
Minimum		0.12	0.17	7.86	1.13	0.28	0.20	0.27	0.65	0.11
Maximum		2.31	2.85	65.51	19.15	4.77	3.31	4.50	11.03	1.88
Mean		1.23	1.44	32.37	9.86	2.46	1.71	2.43	5.74	0.99
Standard		0.76	0.96	24.42	6.26	1.56	1.08	1.46	3.56	0.60
Deviation										
(± represents	(± represents counting error)									

(± represents counting error)

The activity concentrations of ²²⁶Ra varied from 0.12±0.02 Bq l⁻¹ in Oluwa river water sample (Opa) to 2.31±0.35 Bq l⁻¹ in Akeun river water sample (Ikoya 1) with a mean (± standard deviation (SD)) of 1.23±0.76 Bq l⁻¹. Radium-228 activity concentration varied from 0.17±0.03 Bq l⁻¹ in Oluwa river water sample (Opa) to 2.85±0.47 Bq l⁻¹ in Ufara river water sample (Oloto 2) with a mean (±SD) value of 1.44±0.96 Bq l⁻¹. That of ⁴⁰K ranged from 7.86±1.18 Bq l⁻¹ in Oluwa river water sample (Opa) to 65.51±2.43 Bq l⁻¹ in Akeun river water sample (Ikoya 2) with a mean (±SD) value of 32.37±24.42 Bq l⁻¹. In 50% of the water samples, activity concentration of radionuclides is of the

order $^{226} Ra < ^{228} Ra < ^{40} K$ while in the remaining 50%, it is of order $^{228} Ra < ^{238} U < ^{40} K$. River Oluwa water sample (Opa) exhibits the minimum activity concentration for all the natural radionuclides while river Akeun water sample (Ikoya 1) exhibits the maximum activity concentration for $^{226} Ra$, $^{228} Ra$ maximum activity concentration was exhibited by Ufara river water sample (Oloto 2) and river Akeun water sample (Ikoya 2) presents the maximum activity concentration of 65.51 Bq l-¹ for $^{40} K$.

Table 3 displays the minimum, maximum and mean (\pm SD) of the activity concentrations of the radionuclides for the 5 rivers.

Table 3: Minimum, Maximum and Mean Activity Concentrations of 226Ra, 228Ra and 40K in the River Water Samples

River	Statistics	Activity c	Activity concentration (Bq l-1)			
		²²⁶ Ra	²²⁸ Ra	⁴⁰ K		
Oluwa	Minimum	0.12	0.17	7.86		
	Maximum	0.58	0.52	21.34		
	Mean	0.35	0.35	14.60		
	Standard Deviation	0.23	0.18	6.74		
	Standard Error	0.16	0.12	4.77		
Akeun	Minimum	1.77	1.78	52.76		
	Maximum	2.31	2.77	65.51		
	Mean	2.04	2.28	59.14		
	Standard Deviation	0.27	0.50	6.38		
	Standard Error	0.19	0.35	4.51		
Ufara	Minimum	0.82	2.29	56.38		
	Maximum	2.18	2.85	64.90		
	Mean	1.50	2.57	60.64		
	Standard Deviation	0.68	0.28	4.26		
	Standard Error	0.48	0.20	3.01		
Oha	Minimum	0.58	0.44	7.95		
	Maximum	2.03	1.25	10.91		
	Mean	1.31	0.85	9.43		
	Standard Deviation	0.73	0.41	1.48		
	Standard Error	0.51	0.29	1.05		
Otu	Minimum	1.09	1.14	14.26		
	Maximum	1.31	1.22	21.81		
	Mean	1.20	1.18	18.04		
	Standard Deviation	0.11	0.04	3.78		
	Standard Error	0.08	0.03	2.67		

River Akeun gave the highest mean activity concentration for ²²⁶Ra (2.04 Bq l-1) while River Oluwa exhibits its lowest activity concentration (0.35 Bq l⁻¹). The highest mean activity concentration for 228Ra (2.50 Bq l⁻¹) was found in River Ufara while River Oluwa showed its lowest activity concentration (0.35 Bq l⁻¹).

The calculated total annual effective doses for all river water samples varied from 1.13 to 19.15 mSv y-1. for age group $0-1\ y$, 0.28 to 4.77 mSv y^1 for 1-2 y age group, 0.20 to 3.31 mSv y^1 for age group 2-7 y, 0.27 to 4.50 mSv y⁻¹ for 7-12 y age group, 0.65 to 11.03 mSv y^{-1} for age group 12 - 17 y and 0.11 to 1.88mSv y^{-1} for age group > 17 y, with mean values of 9.86, 2.46, 1.71, 2.43, 5.74 and 0.99 mSv y⁻¹ respectively. The mean values show that infants in the age group from birth to 1 y receive highest radiation dose than all other age groups. These values are higher than recommended reference levels of 0.26, 0.20 and 0.1 mSv y^{-1} for infants (0-1 y and 1-2 y age groups), children (2-7 y and 7-12 y age groups) and adults $(12-17 \text{ y and } >17 \text{ y$ y age groups) respectively by the World Health Organization, International Atomic Energy Agency and the United Nations Scientific Committee on Effects of Atomic Radiation [1, 2, 27]. The variations of total annual effective dose values to different age groups are shown in Fig. 1

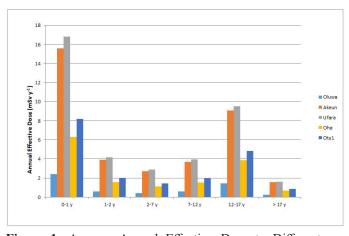


Figure 1: Average Annual Effective Dose to Different age groups from river water consumption

All the river water samples show annual effective doses that exceed the World Health Organization safe limit of 0.1 mSv y⁻¹ from drinking water for all age groups except river Oluwa water for age groups 1-2, 2-7 y, 7-12 y and >17 y [2, 13]. All samples cross the safe limit of annual effective dose of 1.0 mSv y⁻¹ set by the International Commission on Radiological Protection

[28]. Water samples from River Ufara present the highest radiation doses to all age groups while River Oluwa water samples impart the least radiation dose from all radionuclides investigated. The variations of total annual effective dose values in the different rivers are shown in Fig. 2.

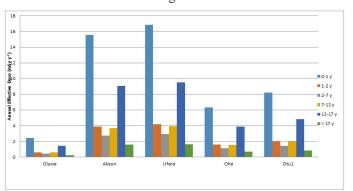


Figure 2: Average Annual Effective Dose from the rivers

The total annual effective doses from various drinking water sources including groundwater, dug well, surface (river) water, sachet, bottled, etc. have been determined at different locations. Annual effective dose values higher than permissible limit in drinking water have been found in some parts of southwestern Nigeria. Achuka reported 5.21 to 14.26 mS y⁻¹, 5.36 – 14.25 $mSv\ y^{-1},\ 3.85\ to\ 10.22\ mSv\ y^{-1},\ 3.82\ to\ 10.13\ mSv\ y^{-1},\ 6.34\ -$ 16.77 mSv y⁻¹ and 3.31 to 8.75 mSv y⁻¹ with a mean value of 9.58, 9.58, 6.87, 6.86, 11.91 and 5.86 mSv y⁻¹ for age groups 0 -1 y, 1-2 y, 2-7 y, 7-12 y, 12-17 y and > 17 y respectivelyin drinking water samples in Ogun State [29]. Similarly Ajayi and Adesida reported annual effective doses in the range 4.73 to 49.13 mSv y^1 , 1.21 to 12.26 mSv y¹, 0.86 to 8.54 mSv y¹, 1.22 to $11.66 \, \text{mSv} \, \text{y}^{-1}, 3.40 \, \text{to} \, 28.98 \, \text{mSv} \, \text{y-1} \, \text{and} \, 0.68 \, \text{to} \, 5.04 \, \text{mSv} \, \text{y}^{-1} \, \text{with}$ mean values of 19.14, 4.87, 3.46, 4.89, 13.36 and 2.61 mSv y⁻¹ for age group 0 - 1 y, 1 - 2 y, 2 - 7 y, 7 - 12 y, 12 - 17 y and > 17y respectively in some sachet drinking water samples produced in Nigeria [30].

In like manner, Ajayi and Owolabi reported annual effective doses in the range 0.05 to 481.60 mSv y⁻¹, 0.02 to 76.84 mSv y⁻¹ and 0.01 to 35.80 mSv y⁻¹ for age groups < 1 y, 2 - 7 y and >17 y respectively in drinking water from private dug wells in Akure. Nigeria [31]. Also Ajayi and Achuka reported annual effective

doses in the range 0.04 to 6.82 mSv y-1, 0.01 to 1.36 mSv y-1 and 0.01 to 1.49 mSv y^{-1} for age groups < 1 y, 2 - 7 y and > 17y respectively in drilled and dug well drinking waters of Ogun State. Nigeria [32]. Nwankwo reported annual effective doses in the range of $0.81 - 1.74 \text{ mSv y}^{-1}$ (with mean value 1.30 mSv y^{-1}) in groundwater for adults in Tanke-Ilorin, Nigeria [17]. Ndontchueng reported annual effective doses in the range of 0.009 to 0.159 mSv y-1 ($\bar{x} = 0.050 \text{ mSv y-1}$), 0.008 to 0.147 mSv y-1 $(\bar{x} = 0.046 \text{ mSy y}^{-1})$ and 0.003 to 0.045 mSy y⁻¹ (0.015 mSy y-1) for infants, children and adults respectively in some mineral bottled water samples produced in Cameroon. All the results show that infants (from birth to 2 y of age) receive greater radiation doses from drinking water than adults (> 17 years of age) [39]. The accumulation of radiotoxic materials like ²²⁶Ra and ²²⁸Ra and their precursors and progenies in growing bones of babies and children can cause bone cancers. Humans of all ages that consume these waters face the risk of some health hazards resulting from significant buildup of radium in their bones and other radiosensitive soft body tissues.

The majority of households in the study area do not have access to pipe borne (treated) water supply, they rely on water from these rivers and streams in the region for drinking. So, water from rivers in the study area is not safe for drinking for those age groups.

Fishing and farming are the major occupation of people of the study area. Due to contamination of water bodies in the oil-producing neighboring communities, some fishes in rivers of those regions flow against the current and migrate to rivers in the study area. Hence, the study area provides majority of the fresh water fishes consumed in Ondo State and other neighboring states. The high natural radionuclide content of the river waters of this region can in turn lead to high concentration of these radionuclides in the fishes and other aquatic creatures gotten from the rivers thus increasing the radiological health risk of people consuming the fishes.

Radiological Risk Due to Radium Ingestion Mass Concentration of Radium (226Ra and 228Ra)

²²⁶Ra and ²²⁸Ra mass concentrations in the different drinking river water samples of part of Bitumen belt of Ondo State, Nigeria are displayed in Table 4 and Table 5 respectively.

Table 4: Radiological and Chemical Toxicity Due to Ingestion of 226Ra in River Waters of the Study Area

River Sample location			Mass concentration x10-6(μg l ⁻¹)	Radiological risk		Chemical risk	
				Mortality risk (x10 ⁻³)	Morbidity risk (x10 ⁻³)	LADD x10 ⁻¹⁴ (μg kg ⁻¹ d ⁻¹)	Hazard Quotient x10 ⁻¹⁴
Oluwa	Opa	0.12	3.29	0.04	0.04	9.39	8.38
	Araromi Ayeka	0.58	15.90	0.19	0.20	45.43	40.56
Akeun	Ikoya 1	2.31	63.33	0.77	0.80	181.01	161.61
	Ikoya 2	1.77	48.54	0.59	0.61	138.02	123.23
Ufara	Oloto 1	0.82	22.52	0.27	0.28	64.13	52.26
	Oloto 2	2.18	59.70	0.72	0.75	171.20	152.86

Oha	Akotogbo 1	2.03	55.61	0.67	0.70	159.66	142.55
	Akotogbo 2	0.58	15.92	0.19	0.20	45.43	40.56
Otu	Iyara 1	1.09	29.81	0.30	0.38	85.31	76.17
	Iyara 2	1.31	35.94	0.44	0.45	102.33	91.36
Minimum		0.12	3.29	0.04	0.04	9.39	8.38
Maximum		2.31	59.71	0.77	0.80	181.01	161.61
Mean		1.23	35.04	0.42	0.44	100.06	89.33
Standard Deviation		0.72	20.9	0.26	0.26	59.82	53.30
Standard Error		0.23	6.61	0.26	0.08	18.92	16.85

Table 5: Radiological and chemical toxicity due to ingestion of 228Ra in river waters of the study area.

River Sample location		Activity con-	Mass con-	Radiological risk		Chemical risk	
		centration (Bq l ⁻¹)	centration x10 ⁻⁸ (μg l ⁻¹)	Mortality risk (x10 ⁻³)	Morbidity risk (x10 ⁻³)	LADD x10 ⁻¹⁶ (μg kg ⁻¹ d ⁻¹)	Hazard Quotient x10 ⁻¹⁴
Oluwa	Opa	0.17	1.69	0.11	0.16	4.82	4.30
	Araromi Ayeka	0.52	5.16	0.49	0.49	14.70	13.13
Akeun	Ikoya 1	1.78	17.63	1.18	1.66	50.41	45.01
	Ikoya 2	2.77	27.50	1.84	2.59	78.50	70.09
Ufara	Oloto 1	2.29	22.72	1.52	2.14	64.91	57.96
	Oloto 2	2.85	28.30	1.89	2.66	80.70	72.05
Oha	Akotogbo 1	1.25	12.41	0.83	1.17	35.41	31.62
	Akotogbo 2	0.44	4.36	0.29	0.41	12.50	11.16
Otu	Iyara 1	1.22	12.10	0.81	1.14	34.63	30.92
	Iyara 2	1.14	11.34	0.76	1.07	32.32	28.86
Minimum	,	0.12	1.69	0.11	0.16	4.82	4.30
Maximum		2.31	28.30	1.89	2.66	80.70	70.09
Mean		1.23	14.30	0.96	1.38	40.90	36.50
Standard Deviation		0.72	9.48	0.63	0.95	27.10	24.20
Standard Err	ror	0.23	3.00	0.20	0.30	8.57	7.65

²²⁶Ra mass concentration varied from 3.29x10⁻⁶ to 59.70x10⁻⁶ μg l⁻¹ with a mean value of 35.04x10⁻⁶ μg l⁻¹ while that of ²²⁸Ra varied from 1.69x10⁻⁸ μg l⁻¹ to 28.30x10⁻⁸ μg l⁻¹ with a mean value of 14.30 μg l⁻¹. The variations may be attributable to differences in local geology of the source of river, geographical location or depth of sample collection. The tables show that all the water samples have ²²⁶Ra and ²²⁸Ra mass concentration below the 20 μg l⁻¹ limit set for Canada 30 μg l⁻¹ limit recommended by the permissible limit set for drinking water in the USA and the 60 μg l⁻¹ limit set for India [13, 14, 24, 32, 33, 34, 35].

Lifetime Cancer Risks

The lifetime cancer risks associated with the ingestion of ²²⁶Ra and ²²⁸Ra radium in the drinking waters of the rivers were evaluated in terms of mortality and morbidity risks. While cancer mortality risk concerns deaths with cancer as the underlying cause in a specified population, morbidity risk is concerned with the amount of cancer within the population. The calculated cancer mortality risk due to the ingestion of ²²⁶Ra are displayed in Table 4, and ranged from 0.04×10^{-3} to 0.77×10^{-3} with mean value of 0.42×10^{-3} whereas cancer morbidity risk ranged from

 $0.04x10^{-3}$ to $0.80x10^{-3}$ with a mean value of $0.44x10^{-3}$. Table 5 shows the calculated cancer mortality due to the ingestion of 228Ra in the water sampled ranged from 0.11x10⁻³ to 1.89x10⁻³ with a mean value of 0.96x10⁻³ while its morbidity risks ranged from 0.16×10^{-3} to 2.66×10^{-3} with a mean value of 1.38×10^{-3} . The mean values of both lifetime cancer risks due to ingestion of both ²²⁶Ra and ²²⁸Ra in the water samples were below the maximum limit of 1.0x10⁻³ specified by USNRC [36]. About 40% of the water samples display ²²⁸Ra cancer mortality risk values above the maximum limit of 1.0x10⁻³ specified USNRC while up to 70% exhibit cancer morbidity risk above 1.0x10⁻³ USNRC limit [36]. The mean carcinogenic risks values of 0.42x10⁻³, and 0.44x10⁻³ (mortality) and 1.38x10⁻³ (morbidity) in this study area are higher than those reported for Odeda Area (1.46x10⁻⁴ and 2.24x10⁻⁴ respectively) by Amakom and Jibiri and Ago-Iwoye (1.09x10⁻⁴ and 1.68x10⁻⁴ respectively) by Alausa in Nigeria [37, 38].

Chemical Toxicity Lifetime Average Daily Dose

The calculated lifetime average daily dose (LADD) due to the

ingestion of ^{226}Ra and ^{228}Ra in the drinking waters displayed in tables 4 and 5 respectively varied from $9.39x10^{-14}$ to $181.01x10^{-14}~\mu g~kg^{-1}~d^{-1}$ with a mean value of $100.06x10^{-14}~\mu g~kg^{-1}~d^{-1}$ for ^{226}Ra while it varied from $4.82x10^{-16}~t0~80.70x10^{-16}~\mu g~kg^{-1}~d^{-1}$ with a mean value of $40.90x10^{-16}~\mu g~kg^{-1}~d^{-1}for$ ^{228}Ra . The mean values of LADD obtained in this study are much lower than the maximum permissible value of $1.0~\mu g~kg^{-1}~d^{-1}$ recommended for drinking water by [13] . The data presented in tables 4 and 5 shows that all of the river water samples have LADD values that are very much lower than the WHO permissible limit.

Hazard Quotient

The hazard quotient (HQ) values estimated using RfD value of 1.12 μ g kg⁻¹ d⁻¹ recommended by WHO [13] ranged from 8.38x10⁻¹⁴ μ g kg⁻¹ d⁻¹ to 161.61x10⁻¹⁴ μ g kg⁻¹ d⁻¹ with a mean value of 89.33x10⁻¹⁴ μ g kg⁻¹ d⁻¹ for ²²⁶Ra and 4.30x10⁻¹⁶ to 70.09x10⁻¹⁶ μ g kg⁻¹ d⁻¹ with a mean value of 36.50x10⁻¹⁶ μ g kg⁻¹ d⁻¹ for ²²⁸Ra All river water samples in the study area showed HQ values much lower than 1.

Conclusion

The results obtained in this study show that the total annual effective doses of radiation from investigated drinking river water samples to all age groups exceed the 0.1 mSv y⁻¹ safe limit set by WHO and the 1.0 mSv y⁻¹ limit set by ICRP for drinking water [13, 20]. They also indicate that the mass concentrations of ²²⁶Ra and ²²⁸Ra in all the samples are below the 30 µg l⁻¹ permissible limit set by WHO [13]. Whereas about 40% of the water samples display 228Ra cancer mortality risk values above the maximum limit of 1.0x10⁻³ specified USNRC, up to 70% exhibit cancer morbidity risk above 1.0x10⁻³ USNRC limit [36]. All the river water samples collected from the study area show lifetime average daily dose and hazard quotient values lower than the acceptable limit of 1.0. Therefore, the radiological toxicity of radium should be a matter of more interest to the population in the study area than its chemical toxicity risks.

References

- United Nations Scientific Committee on the Effects of Atomic Radiation. (2000). Sources and effects of ionizing radiation, ANNEX B, Exposures from natural radiation sources. UNSCEAR 2000 REPORT, New York, 1, 97-99.
- 2. Galal-Gorchev, H., & Ozolins, G. (1993). WHO guidelines for drinking-water quality. Water supply, 11(3), 1-16.
- Brugge D, Buchner, V. (2012). Radium in the Environment: Exposure pathways and health effects. Rev. Environ. Health. 27(1):1-17.
- Amin, S. A., & Jassim, A. A. (2017). Radium Isotopes Levels in Drinking-Water Samples. Engineering and Technology Journal, 35(1 Part B), 46-49.
- Kim, G., Burnett, W. C., Dulaiova, H., Swarzenski, P. W., & Moore, W. S. (2001). Measurement of 224Ra and 226Ra activities in natural waters using a radon-in-air monitor. Environmental science & technology, 35(23), 4680-4683.
- Abbasi, A., & Bashiry, V. (2016). Measurement of radium-226 concentration and dose calculation of drinking water samples in Guilan province of Iran. International Journal of Radiation Research, 14(4), 361-366.

- 7. Abbasi, A., & Mirekhtiary, F. (2019). Lifetime risk assessment of Radium-226 in drinking water samples. International Journal of Radiation Research, 17(1), 163-169.
- Binesh, A., Arabshahi, H., & Pourhabib, Z. (2011). Radioactivity and dose assessment of heavy radioactive pollution, radon and radium from water sources of 3 northern regions in Iran. International Journal of Physical Sciences, 6(35), 7969-7977.
- Maxwell, O., Wagiran, H., Zaidi, E., Joel, E. S., Tenebe, I. T., Oha, I. A., & Onwuka, O. S. (2016). Radiotoxicity risks of radium-226 (226Ra) on groundwater-based drinking at Dawaki, Kuje, Giri and Sabon-Lugbe area of Abuja, North Central Nigeria. Environmental Earth Sciences, 75(14), 1-9.
- National Research Council. Committee on the Biological Effects of Ionizing Radiation (BEIR IV). 1988. Health Risks of Radon and Other Internally Deposited Alpha-Emitters: BEIR IV.276-302.
- 11. International Atomic Energy Agency (IAEA). (2014). The Environmental behavior of radium. Revised Edition. Technical Reports Series. No. 47.
- 12. USEPA. (2000). National Primary Drinking Water Regulations; Radionuclides; Final Rule. Fed. Reg., 65(236), 76708.
- 13. World Health Organization (WHO). (2011). Uranium in Drinking water, Background document for development of WHO guidelines for drinking water quality.
- 14. Health Canada. (1999). Uranium in Drinking Water. Document for Public Comment Prepared by Federal Provincial Subcommittee on Drinking Water.
- 15. Adekoya, O. I. (2014). Primordial radionuclides in potable water from former tin-mining areas with elevated activity. International Journal of Physical Sciences, 9(23), 506-511.
- 16. Agaja, S. A., & Ajisafe, J. I. (2013). Natural radioactivity of surface and ground water samples in coastal communities of Delta State, Nigeria. Int. J. Curr. Res, 5(12), 4057-4061.
- 17. Nwankwo, L. I. (2013). Determination of natural radioactivity in groundwater in Tanke-Ilorin, Nigeria. West African Journal of Applied Ecology, 21(1), 111-120.
- 18. Rahaman, M. A. (1988). Recent advances in the study of the basement complex of Nigeria. Pre Cambrian geology of Nigeria, 11-41.
- Awudu, A. R., Darko, E. O., Schandorf, C., Hayford, E. K., Abekoe, M. K., & Ofori-Danson, P. K. (2010). Determination of activity concentration levels of 238U, 232Th, and 40K in drinking water in a gold mine in Ghana. Health physics, 99(2), S149-S153.
- Clement, C. H., Eckerman, K., Harrison, J., & Menzel, H. G. (2012). Compendium of dose coefficients based on ICRP publication 60. International Commission on Radiological Protection.
- 21. World Health Organization. (2008). Regional and global costs of attaining the water supply and sanitation target (target 10) of the Millennium Development Goals (No. WHO/HSE/AMR/08.01). World Health Organization.
- 22. Singh, L., Kumar, R., Kumar, S., Bajwa, B. S., & Singh, S. (2013). Health risk assessments due to uranium contamination of drinking water in Bathinda region, Punjab state, India. Radioprotection, 48(2), 191-202.
- 23. EPA, U. Environmental Protection Agency (1999), Cancer Risk Coefficients for Environmental Exposure to Radionu-

- clides, Federal Guidance Report N. ^o 13. EPA 402-R-99-001, Office of Radiation and Indoor Air. Online: http://www.epa.gov/radiation/docs/federal/402-r-99-001.
- Kim, Y. S., Park, H. S., Kim, J. Y., Park, S. K., Cho, B. W., Sung, I. H., & Shin, D. C. (2004). Health risk assessment for uranium in Korean groundwater. Journal of environmental radioactivity, 77(1), 77-85.
- World Health Organization, WHO., & World Health Organisation Staff. (2004). Guidelines for drinking-water quality (Vol. 1). World Health Organization.
- Lee, J. S., Chon, H. T., & Kim, K. W. (2005). Human risk assessment of As, Cd, Cu and Zn in the abandoned metal mine site. Environmental geochemistry and health, 27(2), 185-191.
- International Atomic Energy Agency (IAEA). (2002). Specification of Radionuclide content in Commodities Requiring Regulation for Purposes of Radiation Protection. Safety Guide (Draft), Vienna.
- 28. International Commission on Radiological Protection (ICRP). (2000). Protection of the public in situations of prolonged radiation exposure. ICRP Publication 82. Ann ICRP 29(1-2) (Elsevier).
- 29. Achuka, J. A., Usikalu, M. R., & Oyeyemi, K. D. (2017). Radiological risks assessment of Ogun State drinking water. American Journal of Applied Sciences, 14(5), 540-550.
- 30. Ajayi, O. S., & Adesida, G. (2009). Radioactivity in some sachet drinking water samples produced in Nigeria.
- 31. Ajayi, O. S., & Owolabi, T. P. (2008). Determination of natural radioactivity in drinking water in private dug wells in Akure, Southwestern Nigeria. Radiation Protection Dosimetry, 128(4), 477-484.

- 32. Ajayi, O. S., & Achuka, J. (2009). Radioactivity in drilled and dug well drinking water of Ogun state Southwestern Nigeria and consequent dose estimates. Radiation Protection Dosimetry, 135(1), 54-63.
- Gilman, A. P., Villeuve, D. C., Secours, V. E., Yagminas, A. P., Tracy, B. L., Quinn, J. M., ... & Moss, M. A. (1998). Uranyl nitrate: 28-day and 91-day toxicity studies in the Sprague-Dawley rat. Toxicological Sciences, 41(1), 117-128.
- United States Environmental Protection Agency (USEPA).
 (2003). Current Drinking Water Standards. Groundwater and drinking water Protection Agency. Pp. 1-12.
- 35. Board, A. E. R. (2004). Directive for limit on uranium in drinking water. India, Mumbai: AERB.
- 36. Agency (USNRC). (1995). White paper on Risk Harmonization. Final report jointly prepared by NRE and EPA, copy of paper obtained from EPA.
- Amakom, C. M., & Jibiri, N. N. (2010). Chemical and radiological risk assessment of uranium in borehole and well waters in the Odeda Area, Ogun State, Nigeria. International Journal of Physical Sciences, 5(7), 1009-1014.
- 38. Alausa, S. K., Fasunwon, O. O., & Odunaike, K. (2014). Activity and corresponding mass concentrations of 40K and 238U in well waters from Ago-Iwoye, Ogun State, Nigeria. Journal of Natural Sciences Research, 4(12), 1-5.
- 39. Ndontchueng, M. M., Simo, A., Nguelem, E. J. M., Beyala, J. F., & Kryeziu, D. (2013). Preliminary study of natural radioactivity and radiological risk assessment in some mineral bottled water produced in Cameroon. International Journal of Science and Technology, 3(5), 271-276.

Copyright: ©2022 Oladele Samuel Ajayi. This is an open-access article distributed under the terms of the Creative Commons Attribution License, which permits unrestricted use, distribution, and reproduction in any medium, provided the original author and source are credited.